

Ultrasound assisted Fenton-like degradation of dyes using copper doped graphitic carbon nitride

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ABSTRACT

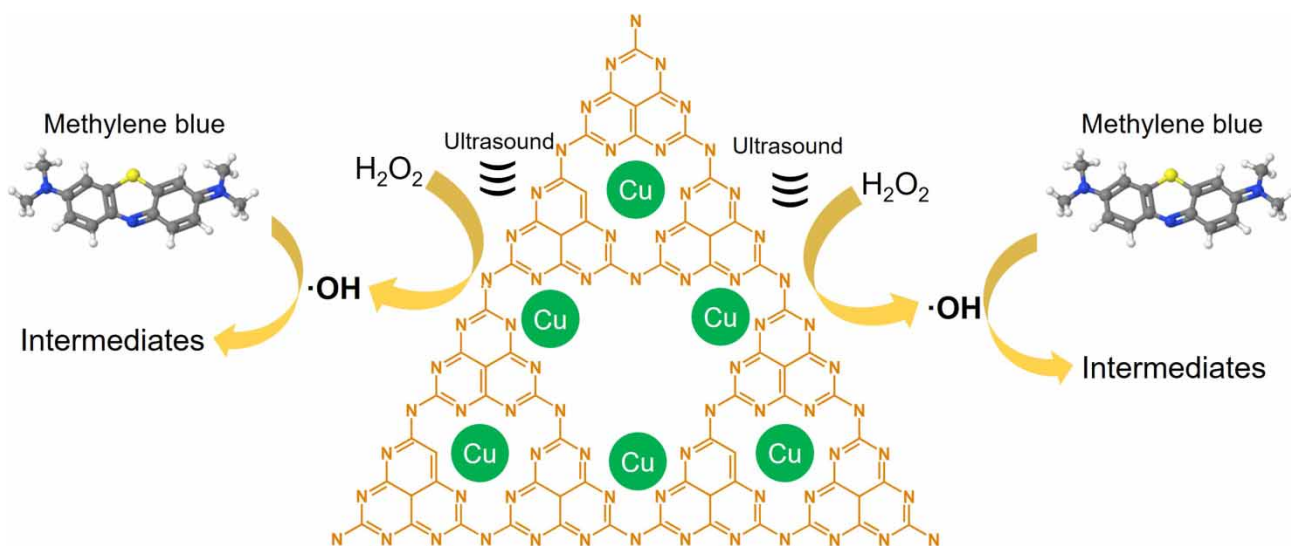
A novel copper doped graphitic carbon nitride (Cu-C₃N₄) was successfully synthesized and used as an effective Fenton-like catalyst. Cu-C₃N₄ was characterized by scanning electron microscopy, surface area analyzer, Fourier transform infrared spectroscopy, X-ray diffractometer, and X-ray photoelectron spectroscopy. Effect of process parameters including catalyst dosage, hydrogen peroxide (H₂O₂) concentration, solution pH, and initial methylene blue (MB) concentration was investigated to evaluate catalytic performance. The pseudo first-order kinetic model was used to describe the catalytic process. The enhancement of MB degradation is observed assisted by ultrasound. MB degradation of 96% is obtained within 30 min in Cu-C₃N₄/H₂O₂/ultrasound system, and the corresponding rate constant is 0.099 min⁻¹. Effective MB degradation is obtained over a broad pH range (3.3–9.9). The catalytic mechanism is examined by ultraviolet-visible spectra, quenching test, and electron spin resonance determination. The dominant mechanism of MB degradation is ascribed to the ultrasonic H₂O₂ activation by Cu-C₃N₄ for hydroxyl radical generation. Cu-C₃N₄ has good reusability and is effective to degrade rhodamine B and acid orange 7. This work not only contributes to the field of wastewater treatment, but also provides insights into the synthesis of Fenton-like catalysts. The results manifest that Cu-C₃N₄ is a promising Fenton-like catalyst for dye degradation in the field of environmental pollution remediation.

Key words: copper doping, dye degradation, Fenton-like reaction, graphitic carbon nitride, hydroxyl radical

HIGHLIGHTS

- A novel Cu doped g-C₃N₄ is easily synthesized.
- Cu species uniformly doped on porous g-C₃N₄ contribute to catalytic ability.
- Ultrasound significantly improves MB degradation in the presence of Cu-C₃N₄/H₂O₂.
- MB degradation is effectively obtained in a broad pH range.
- Plausible mechanism of H₂O₂ activation for •OH generation is revealed for MB degradation.

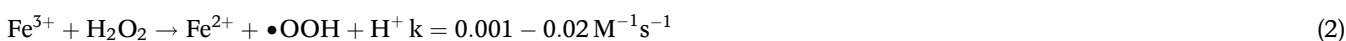
GRAPHICAL ABSTRACT



INTRODUCTION

With rapid industrial development, large amounts of persistent organic pollutants are discharged into the natural environment and cause serious environmental pollution. Approximately 80% of total global wastewater is released into the environment. The wastewater emission poses serious threats to ecosystem and public health (Mian & Liu 2018). Wastewater pollution has been a global challenge in the field of environmental remediation. Dyes are common organic pollutants in wastewater from textile, printing, leather tanning, and cosmetic industries, and they are toxic to ecosystems and the human body (Chanikya *et al.* 2021). There is special concern about removal of toxic dyes from wastewater, which is still challenging (Routoula & Patwardhan 2020).

Conventional techniques, such as chemical coagulation, adsorption, and biological oxidation, generally cause unsatisfactory results for wastewater treatment (Kothai *et al.* 2021). Advanced oxidation processes based on reactive radicals are promising for removing toxic dyes from wastewater due to high oxidizability, complete mineralization, and easy operation. The Fenton reaction is an effective way to generate hydroxyl radicals (•OH) with high oxidation ability ($E = 2.80 \text{ V/NHE}$). Fenton processes are widely applied to wastewater treatment for removing organic pollutants. However, homogeneous Fenton processes suffer from the shortcomings of narrow pH range, producing large amounts of iron sludge, and impractical catalyst recovery (Ma *et al.* 2017). Iron-based Fenton catalysts confront the problem of narrow operation pH (Wang *et al.* 2021a). In this regard, extensive efforts have been conducted to develop effective Fenton-like catalysts operated over a wide pH range (Xin *et al.* 2021). Because of high abundance, low cost, and the striking reactivity towards hydrogen peroxide (H₂O₂), copper-based catalysts have received increasing global attention (Gawande *et al.* 2016). Copper ions are more efficient than iron ions for H₂O₂ activation to generate radical species. As shown in Equations (1)–(4), the rate constant of copper ions towards H₂O₂ is much higher than that of iron ions (Zhang *et al.* 2017; Wang *et al.* 2021a). In addition, H₂O₂ activation by cupric ion occurs over a broad pH range, which is highly attractive for the oxidation of organic pollutants in neutral conditions (Bokare & Choi 2014). Copper-based catalysts, such as copper/carbon composites (Wang *et al.* 2019a) and nanoscale zero-valent copper (Shah *et al.* 2020), have been investigated for Fenton-like degradation of organic pollutants in wastewater. Considering catalytic activity, synthesis process, and catalyst reusability, there is still big research gap for developing effective Fenton-like catalysts:



The synthesis of nanoscale catalysts is an attractive strategy for improving catalytic activity. Metals supported on carbon matrix such as activated carbon, biochar, and graphene have been reported (Wu *et al.* 2020; Wang *et al.* 2021b). Graphitic carbon nitride (g-C₃N₄), a layered material with a 2D π -conjugated polymeric structure, is a promising matrix for loading active metals. The unique physical structure, excellent chemical stability and tunable electronic properties support the photocatalytic activity in environmental applications (Yang *et al.* 2020; Hasija *et al.* 2021). g-C₃N₄ can be easily synthesized through the thermal polymerization of low-cost N-rich precursors such as urea, melamine, dicyandiamide, and thiourea (Jourshabani *et al.* 2020; Nguyen *et al.* 2021). Various modification strategies, such as elemental doping, heterojunction, defects engineering and structural regulation, have been reported to improve the properties for potential applications (Chen *et al.* 2020a). Due to the chemical stability and ligand-field effects, metal doped g-C₃N₄ can be promising candidates for Fenton-like catalysts (Zhu *et al.* 2019).

To solve the problem of developing effective Fenton-like catalysts, a novel copper doped g-C₃N₄ was designed and facilely synthesized. Methylene blue (MB) was chosen as a target pollutant. Ultrasound-assisted MB degradation was evaluated in the presence of H₂O₂. The objectives of this study are (1) to characterize the catalyst, (2) to determine the catalytic performance, and (3) to discuss the plausible catalytic mechanism.

METHODS

Materials

All used chemicals were analytically pure, including hydrogen peroxide (H₂O₂; Sinopharm Chemical Reagent Co., Ltd), copper chloride (CuCl₂·2H₂O; Sinopharm Chemical Reagent Co., Ltd), melamine (C₃H₆N₆; Macklin Reagent Co., Ltd), nitric acid (HNO₃; Sinopharm Chemical Reagent Co., Ltd), sodium hydroxide (NaOH; Sinopharm Chemical Reagent Co., Ltd), isopropyl alcohol (IPA; C₃H₈O; ALADDIN Reagent (Shanghai) Co., Ltd), methylene blue (C₁₆H₁₈ClN₃S; Tianjin Kemiou Chemicals Co., Ltd), rhodamine B (C₂₈H₃₁ClN₂O₅; Tianjin Guangfu Fine Chemical Research Institute), and acid orange 7 (C₁₆H₁₁N₂NaO₄S; ALADDIN Reagent (Shanghai) Co., Ltd). Ultrapure water from a Millipore Milli-Q water purification system was used in experiments.

Catalyst synthesis

The synthesis of catalysts was conducted by metal ion blending and carbonization as shown in Fig. S1 (Supporting Information). First, melamine (2 g) and copper chloride (1.62 g) were dissolved in 10 mL methanol to obtain a uniform solution. Methanol was the chosen solvent because both melamine and copper chloride can be dissolved in it. After removing methanol by stirring at 60 °C, the solid product with uniformly mixed melamine and copper was obtained. Finally, the solid product was calcined in a muffle furnace with limited oxygen. The temperature was elevated to 550 °C with a ramping rate of 5 °C/min and maintained at 550 °C for 120 min. The calcined product was rinsed using ethanol solution (50%) to remove residual substances or by-products of carbonization. After drying at 100 °C for 8 h, the obtained copper doped carbon nitride (Cu-C₃N₄) was used as a catalyst.

Characterizations

A scanning electron microscopy (SEM) equipped with energy dispersive X-ray spectroscopy (EDS; FEI Quanta FEG 250) was used to analyze the morphology and surface elements. The N₂ adsorption-desorption isotherm was determined at 77 K to analyze the surface area using a surface area analyzer (Micromeritics, ASAP 2020, USA). Fourier transform infrared spectroscopy (FTIR; Frontier, PerkinElmer, USA) was used to record the FTIR spectrum. X-ray diffractometer (XRD; Ultimate IV, Rigaku Co., Japan) was employed to determine the chemical phases. X-ray photoelectron spectroscopy (XPS; Thermo ESCALAB 250XI, Thermo Scientific, USA) was used to examine the surface elements. An ultraviolet-visible spectrophotometer (UV-vis; UV-2600, Shimadzu, Japan) was employed to measure the absorbance of characteristic peak and UV-vis spectra of dye solution. A JES FA200 spectrometer was used to detect the electron spin resonance (ESR) signals using 5,5-dimethyl-1-pyrroline-N-oxide (DMPO) as the radical-capture reagent.

Catalytic experiments

The research design of this work is schematically shown in Fig. S2. Batch experiments were conducted in 150 mL conical flask at room temperature. The conical flask containing 50 mL MB solution was put in an ultrasonic cleaner (40 kHz, 150 W; XM-3200UVF, Xiaomei Instrument Limited Company, China). Experimental parameters were investigated including catalyst (0–0.6 g/L), H₂O₂ (0–40 mM), pH (3.3–9.9), and MB concentration (10–30 mg/L). Quenching experiments were

conducted using IPA as radical scavenger. The reusability of the catalyst was studied for MB degradation through five cycles. Catalytic degradation of different dyes was also evaluated. During the experiments, a solution sample was taken out, centrifugally separated, and immediately determined using an UV-vis spectrophotometer. The calculation of MB removal is described in Supporting Information. The pseudo first-order model has been widely used for kinetic analysis. Reaction rate was examined by the kinetic model (described in Supporting Information).

RESULTS AND DISCUSSION

Characterizations

The morphology and elemental distribution of Cu-C₃N₄ are characterized by SEM-EDS. The g-C₃N₄ is a layered massive particle (Hu *et al.* 2019a). As shown in Figure 1, Cu-C₃N₄ shows a sheet-like structure similar to that of g-C₃N₄. Aggregated nanosized particles without well-defined morphology were observed, which may be the doped copper oxides. Cu-C₃N₄ also manifests a porous morphology with intersect channels, which resulted from the release of gaseous products in the heating process (Ma *et al.* 2017; Qin *et al.* 2019). Fig. S3 manifests the major elements of Cu, C, N, and O. Elemental mapping displays the uniform distribution of C, N, O, and Cu in the material, confirming that Cu is doped on g-C₃N₄ (Hu *et al.* 2019a).

As shown in Figure 2(a), the N₂ adsorption-desorption isotherm of Cu-C₃N₄ shows a type IV isotherm with an H3 hysteresis loop according to the IUPAC classification, which suggests the slit-shaped pores resulted from aggregated particles (Ma *et al.* 2017; Jiang *et al.* 2019). The surface area and pore parameters are listed in Table S1 (Supporting Information). The BET surface area is 7.18 m²/g. The pore size distribution plotted by the BJH method displays a bimodal pore distribution ranging from 2 to 10 nm. The average pore diameter is 12 nm.

Figure 2(b) shows the FTIR spectrum of Cu-C₃N₄. The peaks at 3,508–3,423 cm⁻¹ are related to the stretching vibration of the N-H bond and -OH bond (Wang *et al.* 2019b). The characteristic peaks in the region ranging from 1,132 to 1,624 cm⁻¹ (1,132, 1,292, 1,403, and 1,624 cm⁻¹) corresponded to the trigonal C-N(-C)-C or the bridging C-NH-C units (Yao *et al.* 2015). The peak associated with the breathing vibration of the s-triazine units appears at 781 cm⁻¹, lower than that of g-C₃N₄ (at around 807 cm⁻¹) (Muniandy *et al.* 2017). The peak at 549 cm⁻¹ may be ascribed to the stretching vibration of the Cu-N bond, implying that copper species are chemically coordinated to g-C₃N₄ (Dong *et al.* 2017).

The crystallinity of Cu-C₃N₄ was examined by XRD analysis (Figure 2(c)). The diffraction peak at 2-theta of 27.7° is ascribed to the (002) plane for graphitic materials (PDF#87-1526), suggesting the presence of g-C₃N₄ structure (Wang *et al.* 2019b). The diffraction peaks at 2-theta of 32.4°, 35.4°, 38.6°, 48.8°, 53.3°, 58.2°, 61.5°, 66.2°, and 67.9°, which are indexed to the (1 1 0), (1 1 -1), (1 1 1), (2 0 -2), (2 0 0), (2 0 2), (1 1 -3), (3 1 -1), and (1 1 3) crystal planes of CuO (PDF#48-1548) (Feng *et al.* 2019). The peaks at 2-theta of 36.4° and 41.2° are associated with the (1 1 1) and (2 0 0) crystal planes of Cu₂O (PDF#65-3288) (Omran & Nezamzadeh-Ejhieh 2020). The diffraction peaks at 2-theta of 43.2°, 50.4°, and 74.8° are attributed to the (1 1 1), (2 0 0), and (2 2 0) crystal planes of Cu (PDF#65-9026) (Zhang *et al.* 2020a).

The signals of C 1s, N 1s, O 1s, and Cu 2p are observed at around 285 eV, 399 eV, 530 eV, and 932 eV in the XPS spectra of Cu-C₃N₄ (Figure 2(d)). As shown in Figure 2(e), the high-resolution XPS spectra of C 1s manifest the peaks at 284.8 eV (C-C), 286.2 eV (C-O), and 288.3 eV (N-C = N) (Zhang *et al.* 2020b; Song *et al.* 2021). Figure 2(f) displays the high-resolution N 1s spectra. The peaks at 398.9 eV, 399.6 eV and 400.7 eV are assigned to the sp²-hybridized pyridine nitrogen (C-N = C), tertiary pyrrolic nitrogen (N-C₃), and C-N-H groups (Li *et al.* 2018a). High-resolution O 1s spectra show peaks at 529.9 eV and 531.5 eV (Figure 2(g)), which are associated to the oxygen coordinated with copper (Cu-O) and the hydroxyl groups (-OH) (Kumar *et al.* 2014). Figure 2(h) shows the high-resolution XPS spectra of Cu 2p. The binding energy peaks at 932.9 eV and 952.8 eV correspond to Cu⁺ or Cu⁰ (Song *et al.* 2021). The peaks at 935.1 eV and 954.9 eV are attributed to the presence of Cu²⁺ (Liu & Hensen 2013). The peak at 570.1 eV in the Cu LMM spectrum confirms the presence of Cu⁺. In addition, the shakeup satellite peaks are observed at 942.5 eV and 962.5 eV (Liu & Hensen 2013). The peak fittings of high-resolution Cu 2p spectra reveal the existence of Cu²⁺, Cu⁺, and Cu⁰ in Cu-C₃N₄, which agrees with XRD analysis. Combining the FTIR, XRD, and XPS analysis, it can be concluded that copper is chemically incorporated into the g-C₃N₄ structure, probably embedded in the cavities of g-C₃N₄, or formed copper porphyrins and phthalocyanines (Dong *et al.* 2017).

Effect of catalyst dosage

It is reported that ultrasound is effective to enhance Fenton-like reactions (Wang *et al.* 2020a). Fig. S4 shows the comparison of MB degradation under ultrasound and oscillation. MB removal increases progressively with increasing time. The removal is approximately 80% at 90 min under oscillation, while it reaches above 90% at 40 min assisted by ultrasound. The pseudo

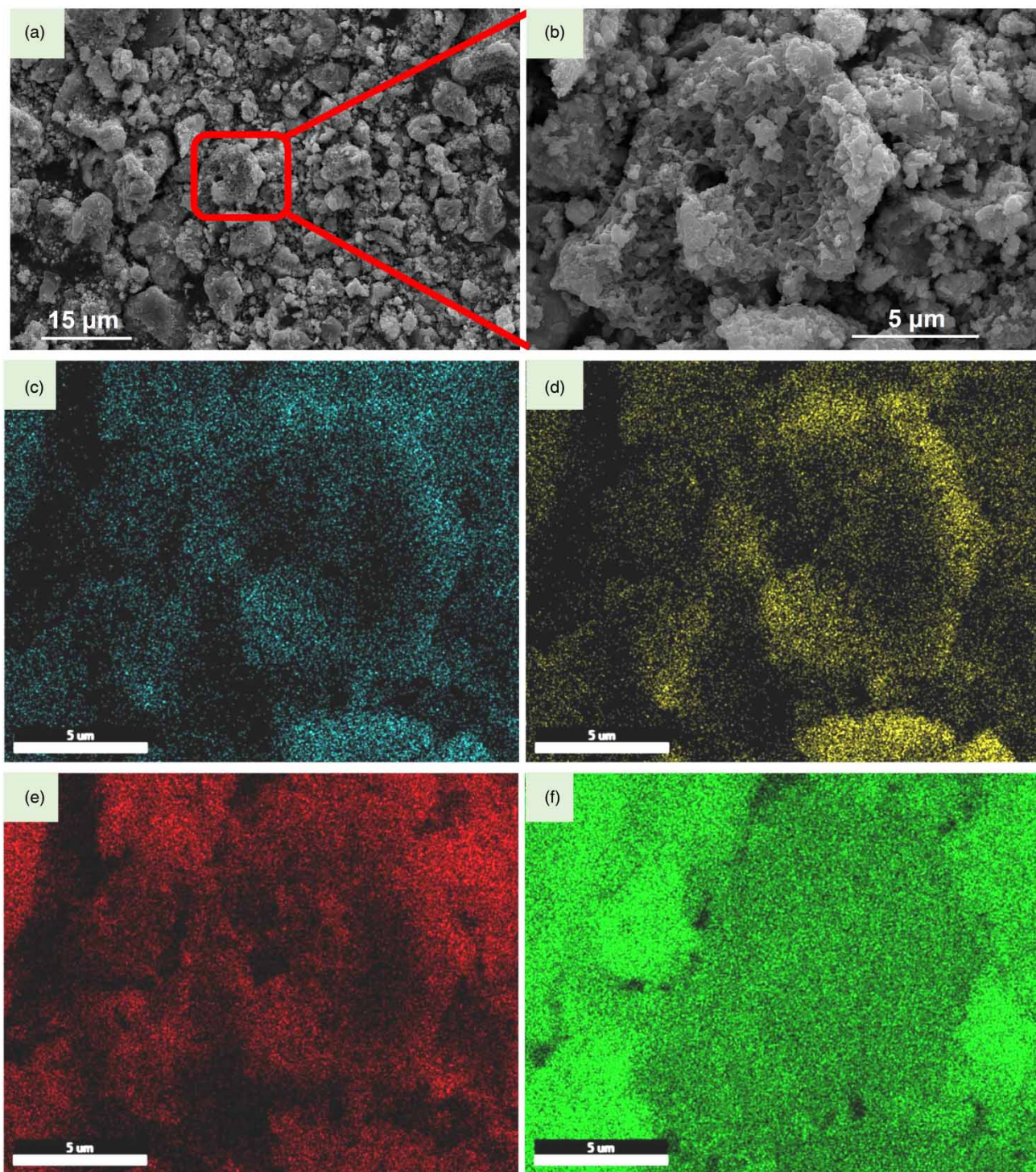


Figure 1 | (a) and (b) Micro-morphology of $\text{Cu-C}_3\text{N}_4$, and elemental distribution: (c) C, (d) N, (e) O, and (f) Cu.

first-order model has been extensively reported for studying the kinetics of Fenton-like processes (Wang *et al.* 2021b). Linear fitting of the model offers the apparent rate constant of pseudo first-order kinetics. The rate constant is significantly improved from 0.019 to 0.048 min^{-1} with the assistance of ultrasound (Fig. S5). The catalytic system of $\text{Cu-C}_3\text{N}_4/\text{H}_2\text{O}_2/\text{ultrasound}$ was used for subsequent degradation experiments.

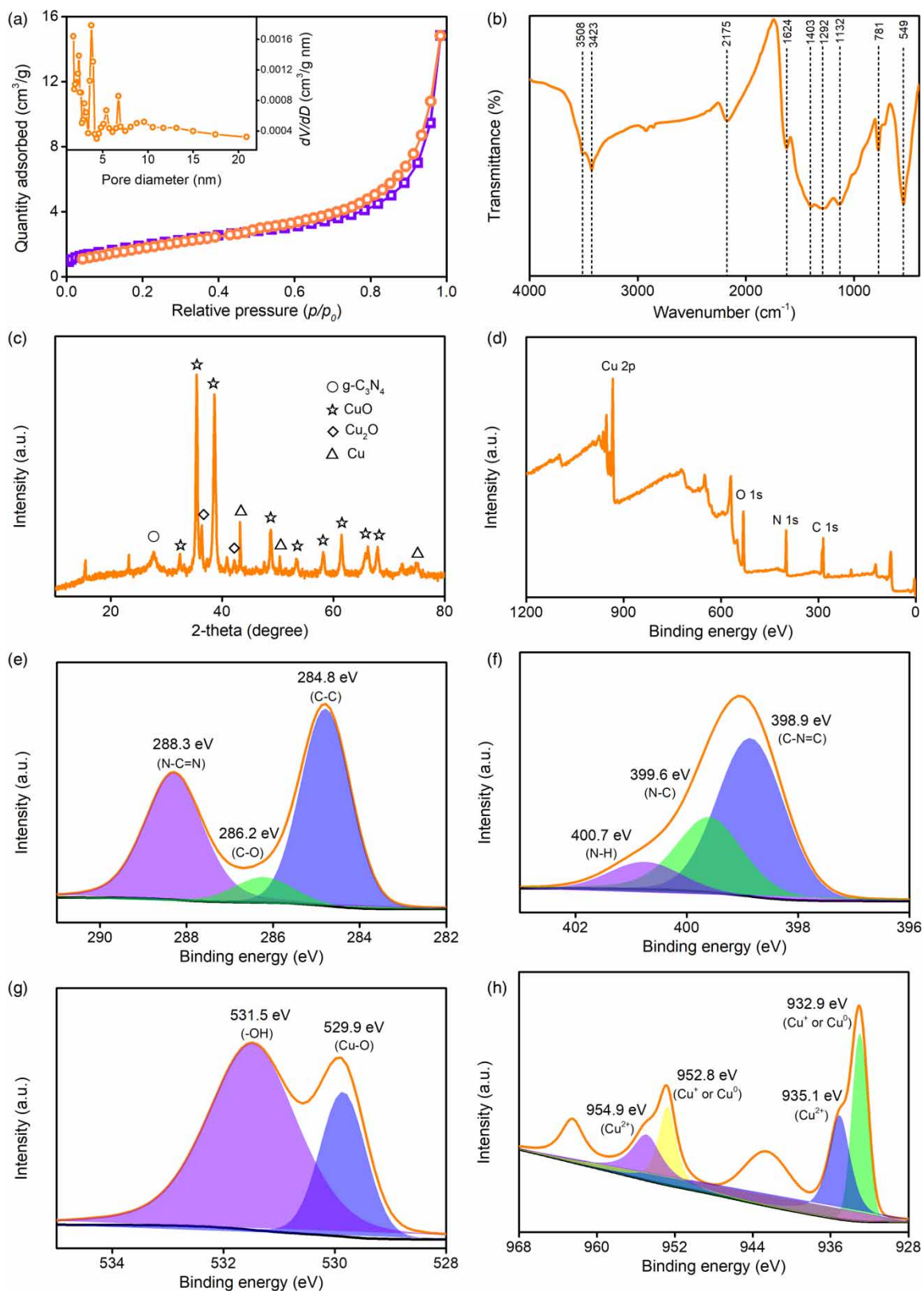
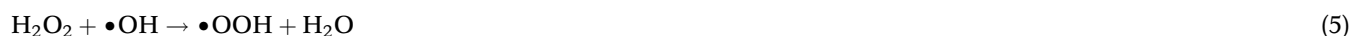


Figure 2 | (a) N_2 adsorption-desorption isotherm, (b) FTIR spectrum, (c) XRD pattern, (d) full XPS spectra, (e) high-resolution C 1s spectra, (f) high-resolution N 1s spectra, (g) high-resolution O 1s spectra, and (h) high-resolution Cu 2p spectra of $Cu-C_3N_4$.

The proper dosage of catalyst needs to be determined since it is closely related to operation cost for application. Figure 3(a) shows MB removal in the presence of different Cu-C₃N₄ dosage. MB removal is notably improved when increasing Cu-C₃N₄ from 0 to 0.1 g/L. When Cu-C₃N₄ is increased from 0.1 to 0.4 g/L, the removal is slightly enhanced. MB removal is nearly unchanged with further increasing Cu-C₃N₄ dosage. The apparent rate constant is 0.004 min⁻¹, 0.029 min⁻¹, 0.042 min⁻¹, 0.081 min⁻¹, and 0.087 min⁻¹, respectively (Figure 3(b)). The improved MB degradation associated with catalyst dosage was reported by Qin *et al.* (2018). The improvement of MB removal with increasing Cu-C₃N₄ is due to introducing more active sites, benefiting the generation of radical species that promote the degradation reactions (Wang *et al.* 2021b). MB removal is not further enhanced with excessive catalyst, probably because of particle agglomeration of the catalyst and undesirable scavenging radicals (Qin *et al.* 2018). 0.4 g/L of Cu-C₃N₄ was selected as optimal dosage for MB removal.

Effect of H₂O₂ concentration

Figure 4(a) displays MB removal in the presence of different H₂O₂ concentration. MB removal is ineffective in the absence of H₂O₂ (Fig. S6), implying the poor adsorption ability of Cu-C₃N₄ and minor effect of ultrasound (Wang *et al.* 2020a, 2021b). MB removal is sharply promoted in the presence of 10 mM H₂O₂. The removal is further improved by adding 20 mM H₂O₂. Further increasing H₂O₂ levels results in slight improvement in MB removal. The effect of H₂O₂ concentration can be verified by the apparent rate constant (Figure 4(b)). Similar results have been reported by Chen *et al.* (2020b) and Xin *et al.* (2021). The positive effect of H₂O₂ is related to the accelerated generation of reactive radicals that are responsible for MB degradation. In the presence of excessive H₂O₂, MB removal is minorly improved. This can be attributed to the undesirable scavenging reactions (Equations (5) and (6)) (Liu *et al.* 2020b; Wang *et al.* 2021b). Additionally, excessive H₂O₂ may cause the self-decomposition into H₂O and O₂ (Equation (7)), and the radical-radical reaction (Equation (8)) causes additional consumption of hydroxyl radical (Yuan *et al.* 2019; Wang *et al.* 2020b). 20 mM of H₂O₂ was selected for MB removal:



Effect of solution pH

Solution pH is a key parameter for the Fenton-like process, and it influences the oxidation potential of reactive radicals, the stability of H₂O₂, and the reactivity of catalysts (Sun *et al.* 2021). The effect of solution pH on MB removal is shown in Figure 5(a). MB degradation remains almost the same over broad pH regions of 3.3–9.9. Although there exists a small difference in the apparent rate constant (Figure 5(b)), effective MB removal is achieved within 40 min under different pH value. This can be ascribed to the wide operation pH of copper ion. Generally, iron-based catalysts suffer from the narrow operation

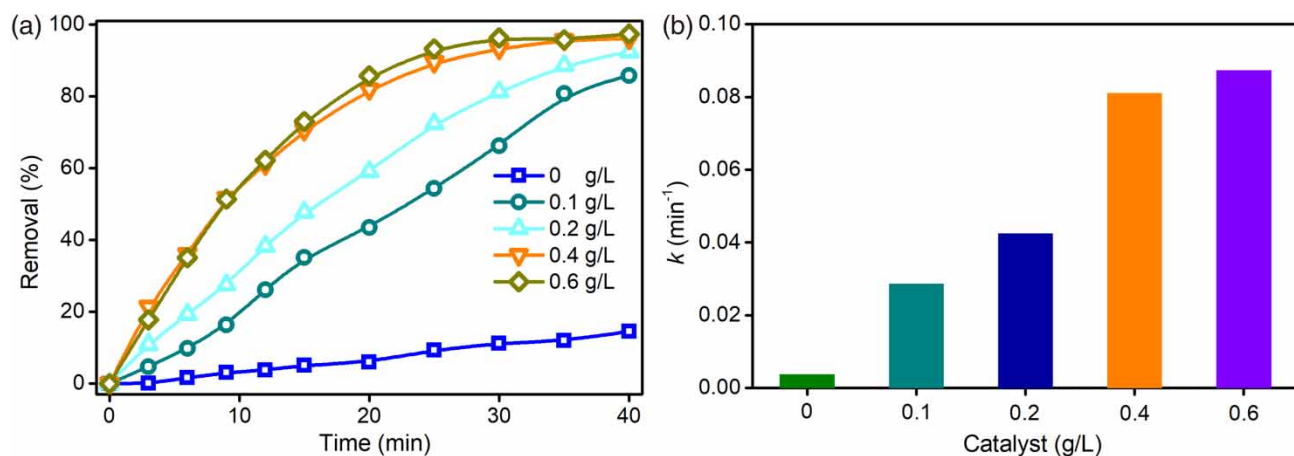


Figure 3 | (a) Effect of Cu-C₃N₄ dosage on MB removal, (b) the apparent rate constant. (Experimental conditions: H₂O₂ 20 mM, solution pH 6.7, MB 10 mg/L, and ultrasound.)

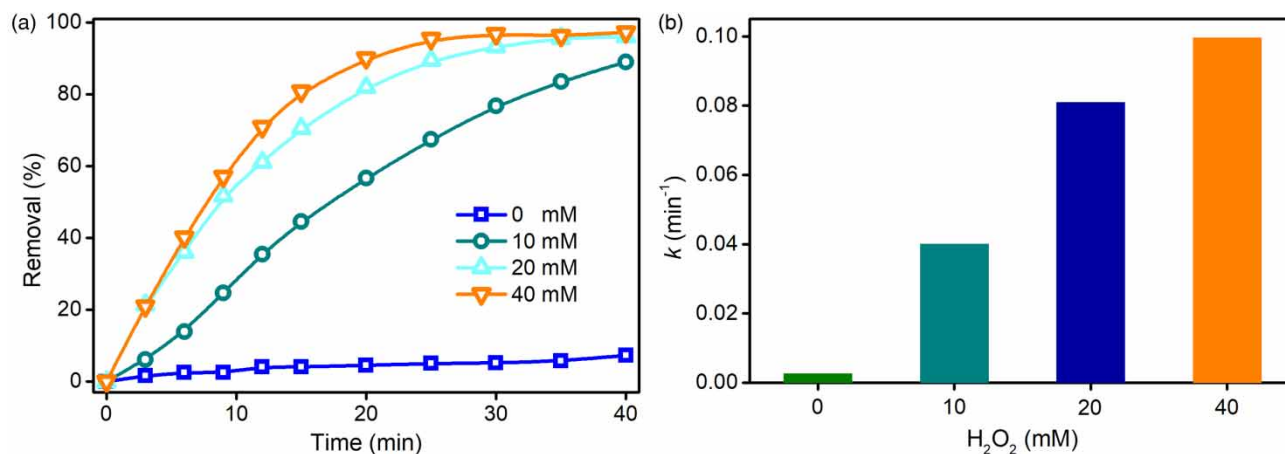


Figure 4 | (a) Effect of H₂O₂ concentration on MB removal, (b) the apparent rate constant. (Experimental conditions: Cu-C₃N₄ 0.4 g/L, solution pH 6.7, MB 10 mg/L, and ultrasound.)

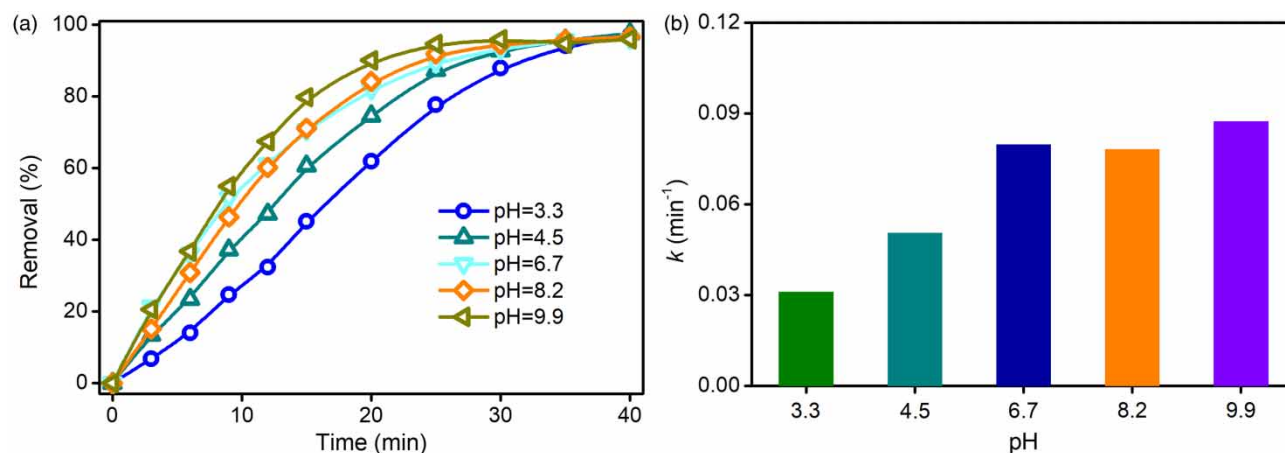


Figure 5 | (a) Effect of solution pH on MB removal, (b) the apparent rate constant. (Experimental conditions: Cu-C₃N₄ 0.4 g/L, H₂O₂ 20 mM, MB 10 mg/L, and ultrasound.)

pH (Yin *et al.* 2020; de Melo Costa-Serge *et al.* 2021). Dye degradation at neutral pH has been highlighted by many researchers (Šuligoj *et al.* 2020; Wang *et al.* 2020b). Effective MB degradation obtained in a broad pH range is superior to the conventional Fenton process and iron-based catalysts.

Effect of MB concentration

The dye concentration in real wastewater varies with time and sources, and thus it is required to evaluate the effect of initial dye concentration. Figure 6(a) displays MB degradation with different initial MB concentration. MB removal increases with increasing reaction time. Effective degradation of MB solution with higher concentration can be obtained by extending reaction time, suggesting the feasibility of MB removal with different concentration. As observed in Figure 6(b), the apparent rate constant is reduced from 0.0810 to 0.0225 min⁻¹ when the initial MB concentration increases from 10 to 30 mg/L. The removal is decreased with an increase in MB concentration. This can be ascribed to the lack of enough reactive radicals (Wang *et al.* 2020b). Additionally, high MB concentration generates more degradation intermediates, which potentially poison catalyst surface (Wang *et al.* 2019c).

Reusability of catalyst

To evaluate the practicability of real application, the reusability of Cu-C₃N₄ was determined in the catalytic system. MB removal is slightly decreased (approximate 7%) for four runs (Figure 7). Obvious decline in MB removal is seen for fifth

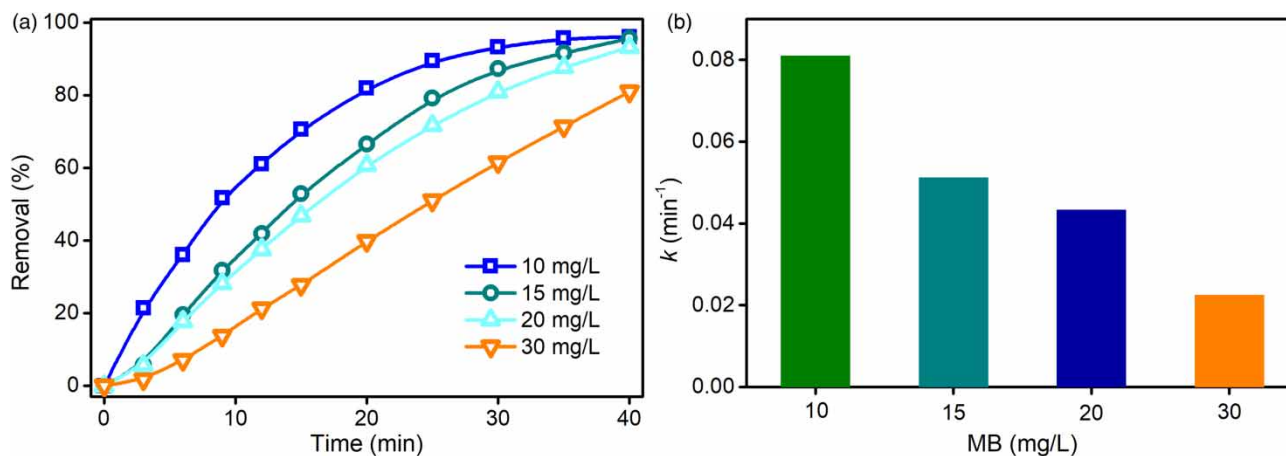


Figure 6 | (a) Effect of initial MB concentration on MB removal, (b) the apparent rate constant. (Experimental conditions: Cu-C₃N₄ 0.4 g/L, H₂O₂ 20 mM, solution pH 6.7, and ultrasound.)

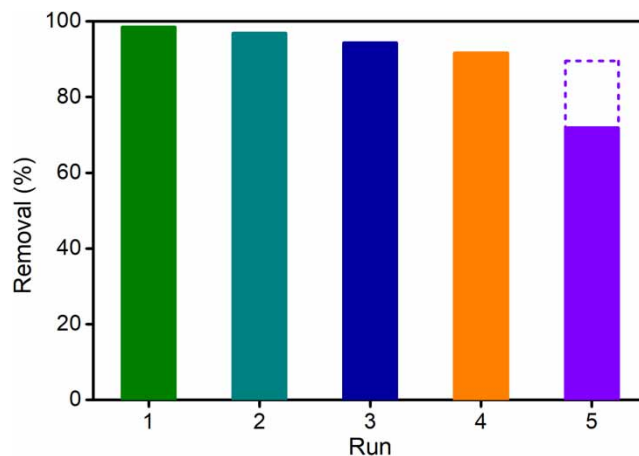


Figure 7 | The reusability of catalyst in Cu-C₃N₄/H₂O₂/ultrasound system. (Experimental conditions: Cu-C₃N₄ 0.4 g/L, H₂O₂ 20 mM, solution pH 6.7, MB 10 mg/L, and ultrasound.)

run, but effective MB removal can be obtained by prolonging reaction time (60 min). The results suggest the good reusability for application. Cu-C₃N₄ as Fenton-like catalyst has outstanding advantages such as high activity, broad operation pH, and good reusability. The process is promising for potential application to wastewater treatment.

Mechanism of catalytic reactions

Figure 8(a) displays the UV-vis spectra of MB solution. The peak at 664 nm is assigned to the chromophore groups (C = S and C = N bonds) in MB molecules (Fig. S7) (Wang *et al.* 2020b). The rapid decrease of the peak at 664 nm is observed with slight shift. This indicates that MB degradation is predominated by chromophore cleavage in the early stage under experimental conditions (He *et al.* 2009). The hypsochromic shift of the characteristic peak at later stage suggests the possible pathway of N-de-ethylation due to the auxochromic property of the N-ethyl group (Mitoraj *et al.* 2018). The degradation intermediates may undergo other reactions including hydroxylation, aromatic ring opening and mineralization (Zhou *et al.* 2015). MB degradation can be verified by the color change of MB solution (Fig. S8). The dark blue color becomes gradually colorless, and complete discoloration is achieved within 40 min. The color change of MB solution is consistent to the time-dependent UV-vis spectra.

MB removal in the presence of Cu-C₃N₄ is negligible within 120 min. The result implies the minor role of adsorption in MB removal, probably due to the small surface area of Cu-C₃N₄. Fenton-like reactions play dominated role in MB removal.

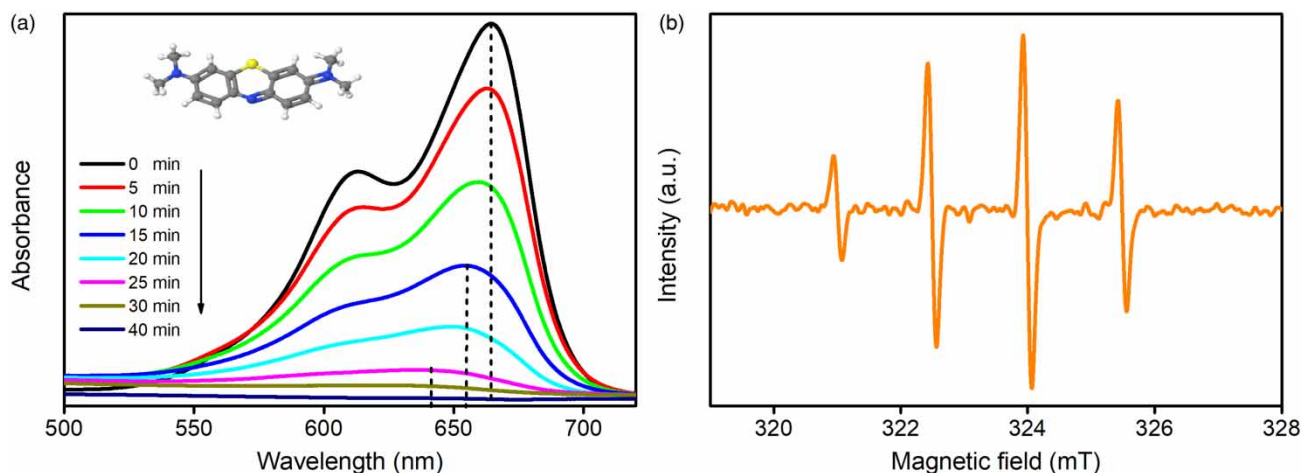


Figure 8 | (a) UV-vis spectra of MB solution associated with different reaction time, (b) ESR spectrum of DMPO-trapped $\bullet\text{OH}$.

Copper species in $\text{Cu-C}_3\text{N}_4$ provide active sites to activate H_2O_2 for the generation of reactive radicals (Liu *et al.* 2021). Cu^0 can be converted into Cu^+ (Equation (9)). Cu^+ reacts with H_2O_2 to generate $\bullet\text{OH}$. Cu^{2+} can be reduced to Cu^+ by reaction with H_2O_2 , realizing the $\text{Cu}^{2+}/\text{Cu}^+$ cycle. In the catalytic system, more radical species are generated assisted by ultrasound. Ultrasound also enables H_2O_2 activation into $\bullet\text{OH}$ radicals (Equations (10)–(12)) (Prakash *et al.* 2020; Wang *et al.* 2020a):



The radicals are responsible for MB degradation. As reported by previous studies, $\bullet\text{OH}$ is the dominant oxidant in Fenton-like processes (Wang *et al.* 2019a; Xin *et al.* 2021). To verify the role of $\bullet\text{OH}$ in the $\text{Cu-C}_3\text{N}_4/\text{H}_2\text{O}_2/\text{ultrasound}$ system, IPA is used as radical scavenger. Fig. S9 displays MB removal, which is lowered in the presence of IPA, and which is attributed to the scavenging effect. The existence of $\bullet\text{OH}$ is further confirmed by ESR spectrum using DMPO as a probe. A strong 4-fold characteristic peak of typical DMPO-trapped $\bullet\text{OH}$ adduct with an intensity ratio of 1:2:2:1 is observed in Figure 8(b) (Xin *et al.* 2021). It should be noted that besides $\bullet\text{OH}$, other radicals cannot be ruled out for MB degradation. Based on above analysis, the plausible catalytic mechanism in $\text{Cu-C}_3\text{N}_4/\text{H}_2\text{O}_2/\text{ultrasound}$ system is schematically shown in Figure 9.

Degradation of different dyes including azo dye (AO7) and triphenylmethane dye (RhB) were further conducted. Figure 10(a) manifests the UV-vis spectra of RhB. The maximum peak at 554 nm is ascribed to the N-ethyl group of RhB (Wang *et al.* 2021b). Effective degradation of RhB is obtained within 30 min as verified by the sharp decline of peak at 554 nm. The pathway of RhB degradation involves chromophore groups and N-de-ethylation (Wang *et al.* 2020a). Similar to MB degradation, the degradation of RhB is dominated by chromophore cleavage, while N-de-ethylation succeeded. Figure 10(b) shows that the typical peak at 484 nm of AO7 significantly declined within 40 min, validating the effective degradation of AO7 in the catalytic system (Wang *et al.* 2019c). Unlike MB degradation, obvious shift of the characteristic peak occurs in the initial stage, suggesting that degradation is not controlled by chromophore cleavage. The results verify that different dyes can be effectively degraded in the $\text{Cu-C}_3\text{N}_4/\text{H}_2\text{O}_2/\text{ultrasound}$ system.

CONCLUSIONS

As verified by SEM, XRD, FTIR, and XPS, $\text{Cu-C}_3\text{N}_4$ contains copper species of Cu, Cu_2O and CuO, which are uniformly doped on the surface of porous $\text{g-C}_3\text{N}_4$. Enhanced MB degradation is observed by ultrasound. MB degradation in the $\text{Cu-C}_3\text{N}_4/\text{H}_2\text{O}_2/\text{ultrasound}$ system is described by pseudo first-order kinetics. MB degradation is influenced by catalyst dosage, H_2O_2 concentration, solution pH, and initial MB concentration. Under conditions of $\text{Cu-C}_3\text{N}_4$ 0.4 g/L, H_2O_2

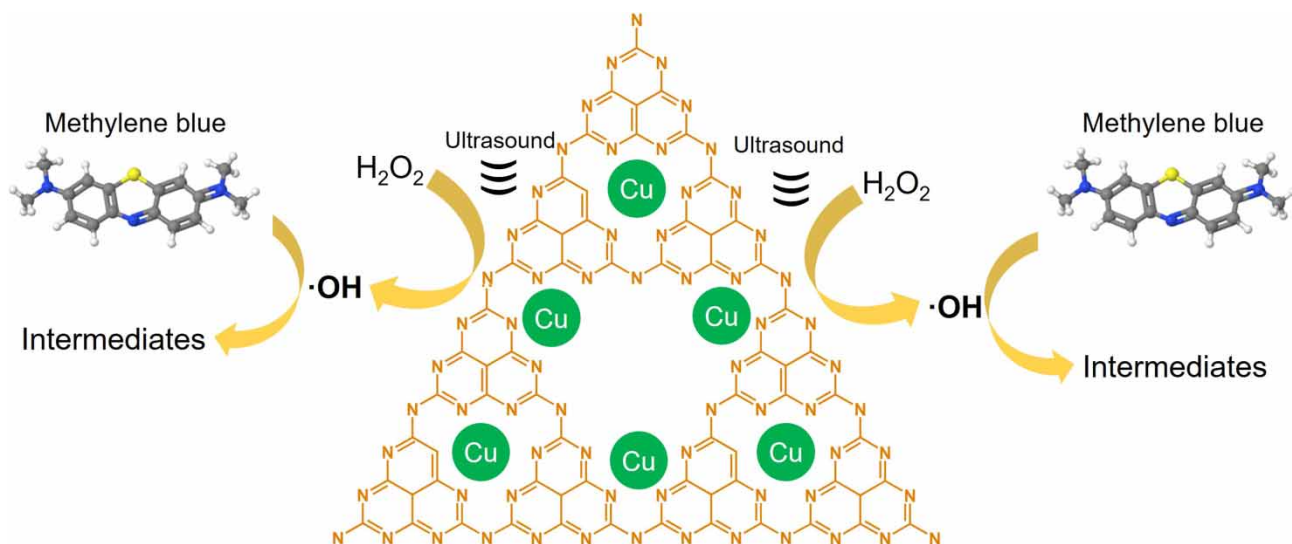


Figure 9 | Schematical scheme of plausible catalytic mechanism in Cu-C₃N₄/H₂O₂/ultrasound system.

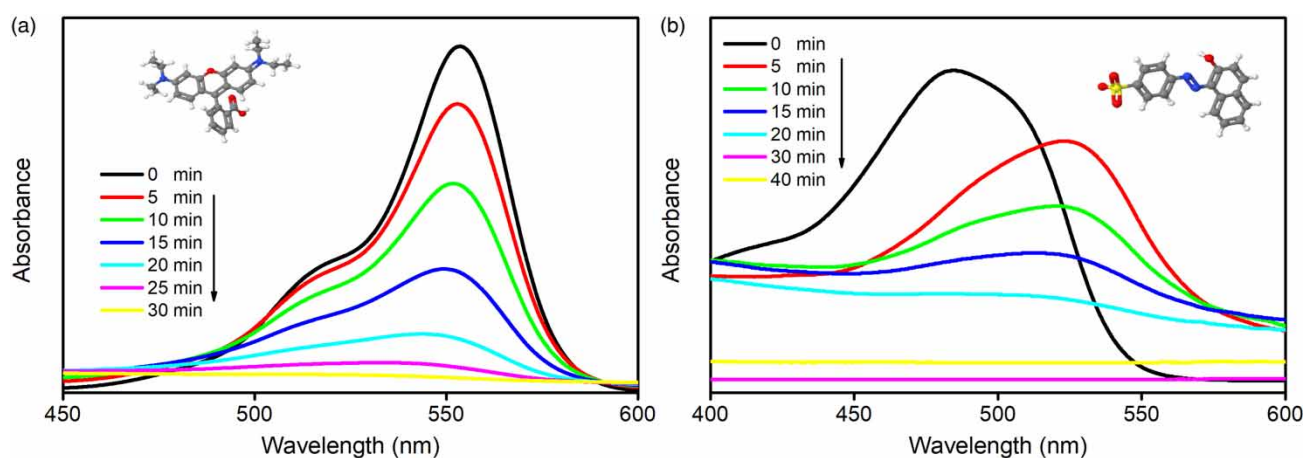


Figure 10 | Time-dependent UV-vis spectra of (a) RhB and (b) AO7 solution.

20 mM, solution pH 6.7, MB 10 mg/L, and ultrasound, MB degradation of 96% is obtained within 30 min, and the corresponding rate constant is 0.099 min^{-1} . Effective MB degradation over a broad pH range (3.3–9.9) is superior to the conventional Fenton process or iron-based catalysts. Cu-C₃N₄ has good reusability as Fenton-like catalysts. Quenching test and ESR spectrum confirmed the existence of $\cdot\text{OH}$ in the catalytic system. Ultrasonically improved H₂O₂ activation by Cu-C₃N₄ for the generation of $\cdot\text{OH}$ radical is the dominant mechanism for MB degradation. The degradation of RhB and AO7 is also effectively achieved. The commercial application of the catalytic system for wastewater treatment requires more efforts. Nevertheless, this work provides insights into designing Fenton-like catalysts and wastewater treatment.

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COMPETING INTERESTS STATEMENT

The authors have no competing interests to declare.

DATA AVAILABILITY STATEMENT

Data cannot be made publicly available; readers should contact the corresponding author for details.

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